

# **Solute Transport Theory & Hydrus applications**

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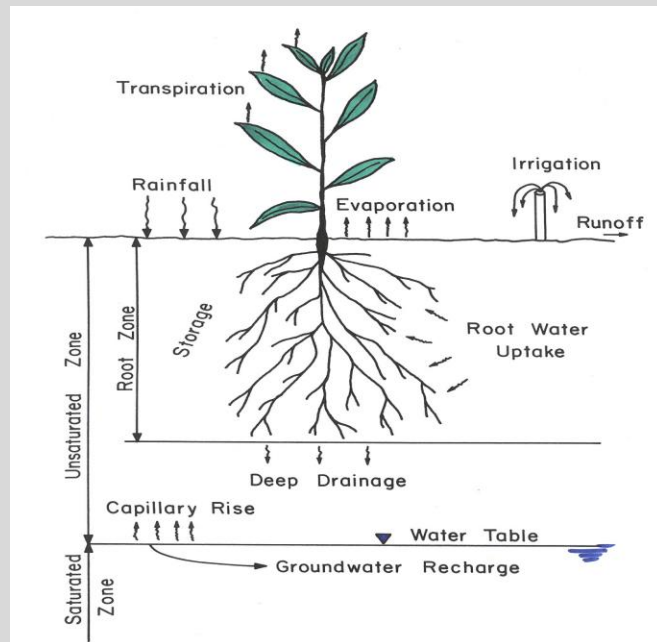
**HYD 210**

**Davis, May - 2011**

## *Outline*

- ◆ **Introduction: Water Flow and Solute Transport**
  
- ◆ **Solute Transport**
  - **Governing transport equations**
  - **Advection, diffusion, dispersion, reactions (ADE)**
  - **Solute Break-Through Curves, Boundary conditions**
  - **Nonequilibrium transport (chemical, physical)**
  - **Decay chains**

## Agricultural Applications



## *Agricultural and Other Contaminants*

- ◆ Fertilizers (nitrogen, phosphorus, potassium)
- ◆ Pesticides (insecticides, fungicides, herbicides)
- ◆ Hormones (estrodial)
- ◆ Pharmaceuticals (antibiotics)
- ◆ Fumigants (methyl bromide)

Salinization and reclamation of sodic soils

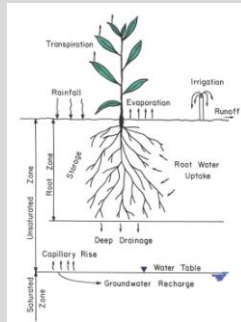
Other pollutants: Heavy metals, radionuclides, hydrocarbons, and many others

# *Increasing Spatial and Time Scales*

**Core or Lab**



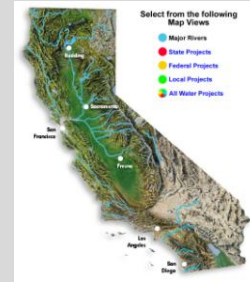
**Plot**



**Field**



**Watershed**



## *Water Flow - Richards Equation*

Governing flow equation for one-dimensional isothermal flow in a variably saturated isotropic rigid porous medium:

$$\frac{\partial \theta(h)}{\partial t} = -\frac{\partial q}{\partial z} - S(h)$$

$$\hat{v} = \frac{q}{\theta}$$

$$\frac{\partial \theta(h)}{\partial t} = \frac{\partial}{\partial z} \left[ K(h) \frac{\partial h}{\partial z} + K(h) \right] - S(h)$$

- $\theta$  - volumetric water content [ $L^3L^{-3}$ ]
- $h$  - pressure head [L]
- $K$  - unsaturated hydraulic conductivity [ $LT^{-1}$ ]
- $z$  - vertical coordinate positive upward [L]
- $t$  - time [T]
- $S$  - root water uptake [ $T^{-1}$ ]

# *Solute Transport*

## *Advection-Dispersion Equation - Simplified*

One-dimensional chemical transport during transient flow in a variably-saturated rigid porous medium

$$\frac{\partial(\theta c)}{\partial t} + \frac{\partial(\rho s)}{\partial t} = \frac{\partial}{\partial z} \left( \theta D \frac{\partial c}{\partial z} - qc \right) - r_c$$

- $c$  - solution concentration [ML<sup>-3</sup>]
- $s$  - adsorbed concentration [MM<sup>-1</sup>]
- $\theta$  - water content [L<sup>3</sup>L<sup>-3</sup>]
- $\rho$  - soil bulk density [ML<sup>-3</sup>]
- $D$  - dispersion coefficient [L<sup>2</sup>T<sup>-1</sup>]
- $q$  - volumetric flux [LT<sup>-1</sup>]
- $r_c$  - rate constant representing reactions [ML<sup>-3</sup>T<sup>-1</sup>]

**HOW DOES IT COME ABOUT ??????**

## *Continuity Equation - Solute*

$$\frac{\partial C_T}{\partial t} + \frac{\partial J_c}{\partial z} + r_c = 0$$

- $C_T$  total concentration of chemical in all forms, as mass of chemical per soil volume [M/L<sup>3</sup>]
- $J_c$  chemical mass flux (mass of chemical flowing per unit area per unit time) [M/L<sup>2</sup> T]
- $r_c$  the rate of loss of mass per unit volume by reactions or other sinks (e.g. plant uptake) [M/L<sup>3</sup> T]

## ***Total Solute Concentration***

$$C_T = \rho_b C_a + \theta C_l + a C_g$$

- $C_a$  adsorbed concentration (mass per unit mass of dry soil), [M/M]  
 $C_l$  dissolved concentration (mass per unit volume of soil solution), [M/L<sup>3</sup>]  
 $C_g$  gaseous concentration (mass of chemical vapor per unit volume of soil air), [M/L<sup>3</sup>]  
 $\rho_b$  bulk density of dry soil, [M/L<sup>3</sup>]  
 $\theta$  volumetric water content, [L<sup>3</sup>/L<sup>3</sup>]  
 $a$  volumetric air content, [L<sup>3</sup>/L<sup>3</sup>]

## ***Total Flux of Chemical in Soil - $J_c$***

$$J_c = J_l + J_g$$

- $J_c$  chemical mass flux (mass of chemical being transported per unit area per unit time) [M/L<sup>2</sup> T]  
 $J_l$  solute liquid flux [M/L<sup>2</sup> T]  
 $J_g$  flux of chemical vapor [M/L<sup>2</sup> T]

**Fick's law for flux in gas phase:**  
**(Gaseous Diffusion only)**

$$J_g = -\xi_g (a) D_g^a \frac{\partial C_g}{\partial z} = -D_g^s \frac{\partial C_g}{\partial z}$$

- $D_g^a$  gas diffusion coefficient in air  
 $D_g^s$  gas diffusion coefficient in soil  
 $\xi_g$  gaseous tortuosity factor

## ***Total Solute Liquid Flux – $J_l$*** [M/L<sup>3</sup> T]

**Total Liquid Flux:**

$$J_l = J_{lc} + J_{lh} + J_{ld}$$

- $J_l$  solute flux [M/L<sup>2</sup> T]
- $J_{lc}$  liquid advective flux [M/L<sup>2</sup> T]
- $J_{lh}$  hydrodynamic dispersion flux [M/L<sup>2</sup> T]
- $J_{ld}$  liquid diffusion flux [M/L<sup>2</sup> T]

## ***Solute Liquid Flux – Diffusion, $J_{ld}$***

**Diffusive flux:**

$$J_{ld} = -\xi_l(\theta) D_l^w \frac{\partial C_l}{\partial z} = -D_l^s \frac{\partial C_l}{\partial z}$$

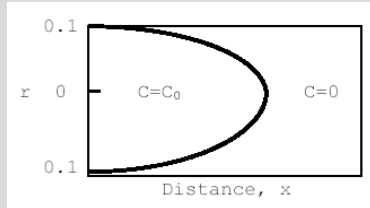
**Tortuosity:**

$$\xi_l(\theta) = \frac{\theta^{10/3}}{\theta_s^2}$$

- $J_{ld}$  liquid diffusion flux [M/L<sup>2</sup> T]
- $\xi_l$  liquid tortuosity factor [-]
- $D_l^w$  binary diffusion coefficient of the solute in water [L<sup>2</sup>/T]
- $D_l^s$  soil liquid diffusion coefficient [L<sup>2</sup>/T]
- $\theta_s$  porosity [-]

# Hydrodynamic Dispersion - $J_{th}$

Fluxes inside a single pore:

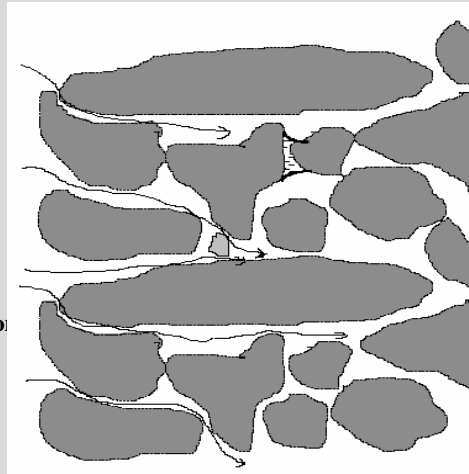


$$v = 2V_0 \left( 1 - \frac{r^2}{a^2} \right)$$

$V_0$  – average flux  
 $a$  – radius of the pore

$$J_{th} = -D_{th} \frac{\partial C_i}{\partial z}$$

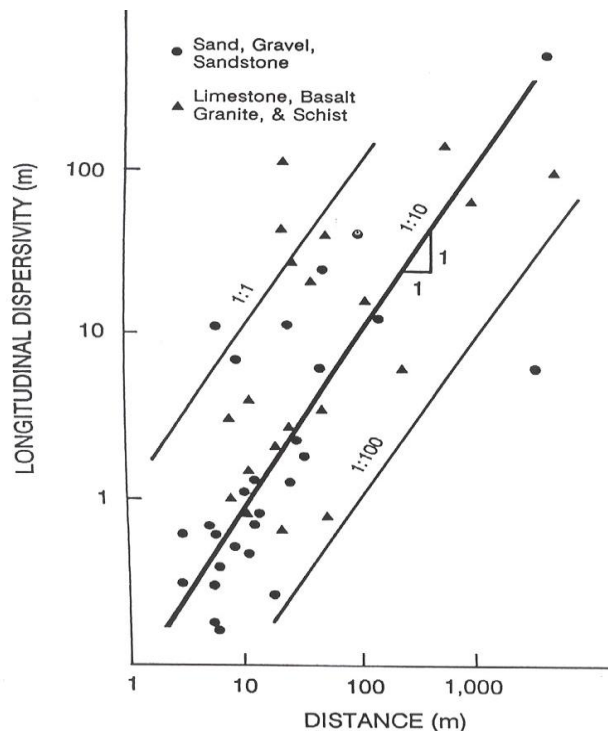
Fluxes inside a system of pores:



Dispersivity as a function of scale

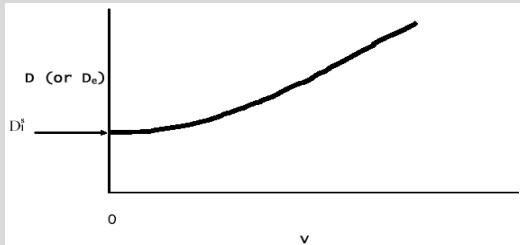
$$D_{h1} = \lambda v$$

$\lambda$  = Dispersivity (L)



# Total Effective Dispersion Coefficient

Many experiments have shown that  $D$  (or  $D_e$ ) increases with the pore-water velocity,  $v$ :



$$D_e = D_{lh} + D_l^s = \lambda v + \frac{\theta^{10/3}}{\phi^2} D_l^w$$

Spreading due to hydrodynamic dispersion is due to complicated flow paths around soil particles, to differences in water velocity within single pores, and to differences in flow velocity between adjacent pores.

# Total Solute Liquid Flux

**Advective Flux:**

$$J_{lc} = J_w C_l$$

**Hydrodynamic Flux:**

$$J_{lh} = -D_{lh} \frac{\partial C_l}{\partial z}$$

**Solute Liquid Flux:**

$$J_l = J_w C_l - D_{lh} \frac{\partial C_l}{\partial z} - D_l^s \frac{\partial C_l}{\partial z} = J_w C_l - D_e \frac{\partial C_l}{\partial z}$$

- $J_l$  solute flux [M/L<sup>2</sup> T]
- $J_{lc}$  liquid advective flux [M/L<sup>2</sup> T]
- $J_{lh}$  hydrodynamic dispersion flux [M/L<sup>2</sup> T]
- $J_{ld}$  liquid diffusion flux [M/L<sup>2</sup> T]
- $J_w$  water flux [L/T]
- $D_{lh}$  hydrodynamic dispersion coefficient [L<sup>2</sup>/T]:  $D_{lh} = \lambda v$
- $\lambda$  dispersivity [L]
- $D_l^s$  soil liquid diffusion coefficient [L<sup>2</sup>/T]
- $D_e$  effective diffusion-dispersion coefficient [L<sup>2</sup>/T]



## ***General Solute Transport Equation***

For the most general case where chemicals are present in the dissolved, sorbed and vapor phases, the chemical transport equation is obtained by combining the continuity equation for the total solute with the vapor and liquid flux densities and the reaction term to produce:

$$\begin{aligned} \frac{\partial C_r}{\partial t} &= -\frac{\partial J_c}{\partial z} - r_c \\ \frac{\partial}{\partial t} (\rho_b C_a + \theta C_l + a C_g) &= \\ &= \frac{\partial}{\partial z} \left( \theta D \frac{\partial C_l}{\partial z} \right) + \frac{\partial}{\partial z} \left( D^s \frac{\partial C_g}{\partial z} \right) - \frac{\partial}{\partial z} (J_w C_l) - (\rho_b r_a + \theta r_l + a r_g) \end{aligned}$$

## ***Chemical Reactions***

$$r_c = \rho_b r_a + \theta r_l + a r_g$$

- $r_c$  rate of loss of mass per unit volume by reactions or other sinks (e.g. plant uptake) [M/L<sup>3</sup>/T]
- $r_a$  reactions in the sorbed phase per unit mass of soil [M/M T]
- $r_l$  reactions in the liquid phase per unit volume of water [M/L<sup>3</sup> T]
- $r_g$  reactions in the gaseous phase per unit volume of air [M/L<sup>3</sup> T]

## Advection-Dispersion Equation (ADE)

When the vapor phase is negligible,  $C_g$  and  $r_g$  are zero, and the ADE becomes:

$$\frac{\partial}{\partial t}(\theta C + \rho_b C_a) = \frac{\partial}{\partial z} \left( \theta D \frac{\partial C_l}{\partial z} \right) - \frac{\partial}{\partial z} (J_w C_l) - (\theta r_l + \rho_b r_a)$$

For linear (first-order) decay such that  $r = \mu c$ :

$$\frac{\partial(\theta C_l + \rho C_a)}{\partial t} = \frac{\partial}{\partial z} \left( \theta D \frac{\partial C_l}{\partial z} \right) - \frac{\partial}{\partial z} (J_w C_l) - \theta \mu_l C_l - \rho \mu_a C_a$$

$\mu$ : First-order transformation rate constant (1/T)  
Can be related to half-life ( $T_{1/2}$ ):

$$\mu = \frac{\ln(2)}{T_{1/2}}$$

## One-Dimensional Equilibrium Transport

General Transport Equation; no production/decay

$$\frac{\partial(\theta c + \rho s)}{\partial t} = \frac{\partial}{\partial x} \left( \theta D \frac{\partial c}{\partial x} \right) - \frac{\partial qc}{\partial x}$$

Equilibrium Sorption ( $s = k c$ )

$$\frac{\partial(\theta R c)}{\partial t} = \frac{\partial}{\partial x} \left( \theta D \frac{\partial c}{\partial x} \right) - \frac{\partial qc}{\partial x}$$

$$R = 1 + \frac{\rho k}{\theta}$$

(Retardation Factor)

Steady-State Flow (homogeneous system; constant  $q, \theta$ )

$$R \frac{\partial c}{\partial t} = D \frac{\partial^2 c}{\partial x^2} - v \frac{\partial c}{\partial x}$$

$$v = \frac{q}{\theta}$$

# Effects of diffusion and advection on transport of solute

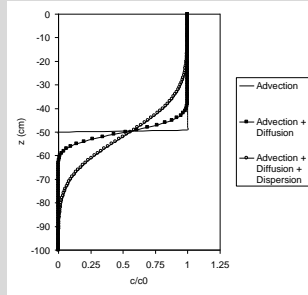
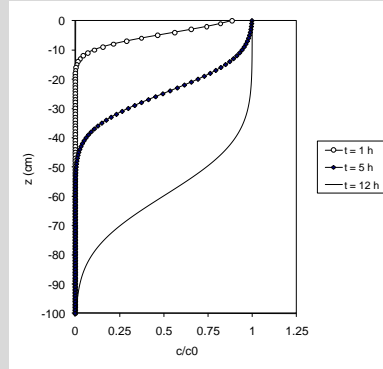


Figure 6.5 Effect of hydrodynamic dispersion and diffusion in dispersing a solute front at  $t = 10$  h with a mean pore water velocity of  $5 \text{ cm h}^{-1}$ .



## Pulse input of solute in soil column ( $v = 5 \text{ cm/hr}$ )

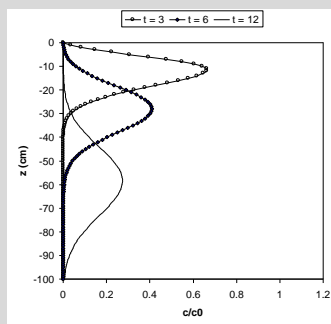


Figure 6.14 Relative concentrations as a function of depth for times of 3, 6, and 12 h predicted by Equation 6.32 for a pulse input with  $v = 5 \text{ cm h}^{-1}$  for  $D_L = 10 \text{ cm}^2 \text{ h}^{-1}$ .

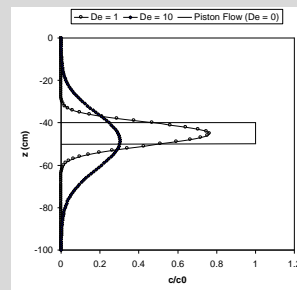


Figure 6.12 Relative concentrations as a function of depth predicted by Equation 6.32 after 10 h for a pulse input of 2 h duration with  $v = 5 \text{ cm h}^{-1}$  and for  $D_L = 1$  and  $10 \text{ cm}^2 \text{ h}^{-1}$  compared to piston flow.

# Effects of solute retardation on transport of solute

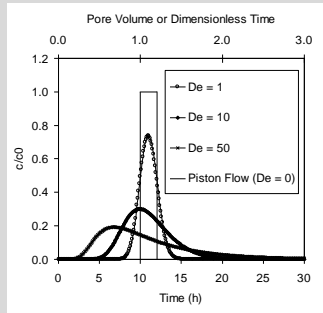


Figure 6.15 BTCs at a depth of 50 cm predicted by Equation 6.32 for a pulse input with  $v = 5 \text{ cm h}^{-1}$  for  $D_e = 1, 10, \text{ and } 50 \text{ cm}^2 \text{ h}^{-1}$  compared to piston flow plotted as a function of time (bottom axis) and as a function of pore volume or dimensionless time (top axis).

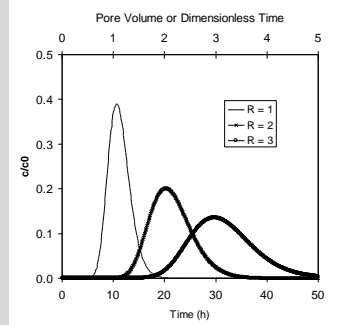


Figure 6.19 BTCs for a depth of 50 cm for a pulse input of 2 h duration with  $v = 5 \text{ cm h}^{-1}$ ,  $D_e = 5 \text{ cm}^2 \text{ h}^{-1}$ , and  $R = 1, 2, \text{ and } 3$  using Equation 6.42 and Equation 6.32 as a function of time (bottom axis) and pore volume or dimensionless time (top axis).

# Effects of solute decay on transport

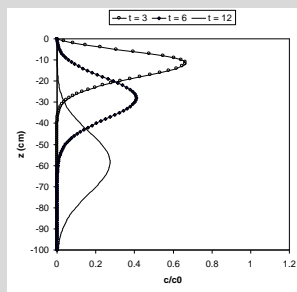


Figure 6.14 Relative concentrations as a function of depth for times of 3, 6, and 12 h predicted by Equation 6.32 for a pulse input with  $v = 5 \text{ cm h}^{-1}$  for  $D_e = 10 \text{ cm}^2 \text{ h}^{-1}$ .

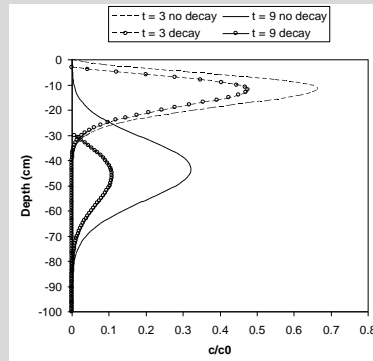
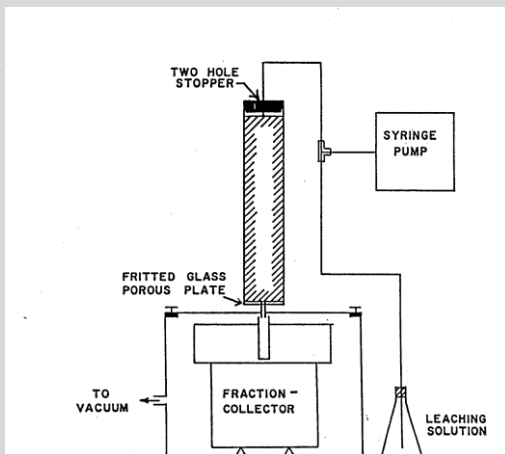


Figure 6.22 Relative concentration as a function of depth with and without decay for a pulse input after 3 and 9 h ( $\mu = 0.1 \text{ h}^{-1}$ ,  $t_0 = 2 \text{ h}$ ,  $v = 5 \text{ cm h}^{-1}$ ,  $D_e = 10 \text{ cm}^2 \text{ h}^{-1}$ ,  $R = 1$ ).

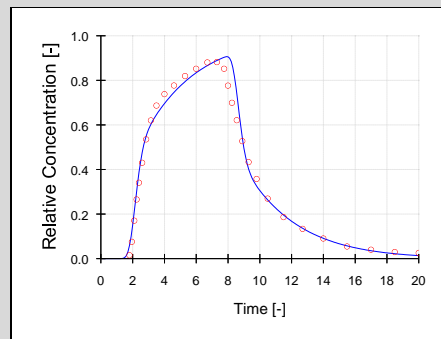
# Outline

- ◆ Introduction: Water Flow and Solute Transport
- ◆ Solute Transport
  - Governing transport equations
  - Advection, diffusion, dispersion, reactions (ADE)
  - Solute Break-Through Curves, Boundary conditions
  - Nonequilibrium transport (chemical, physical)
  - Decay chains

## Soil Column Breakthrough Experiment

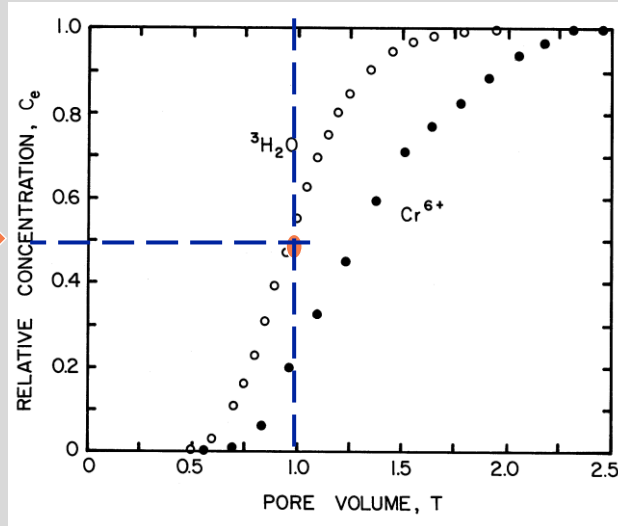


A plot of outflow concentration versus time is called an effluent curve, an outflow curve, or a **"breakthrough" curve** (representing the solute breaking through at the outflow end).



## No sorption (tritium) versus sorption ( $Cr^{6+}$ )

Bulls Eye

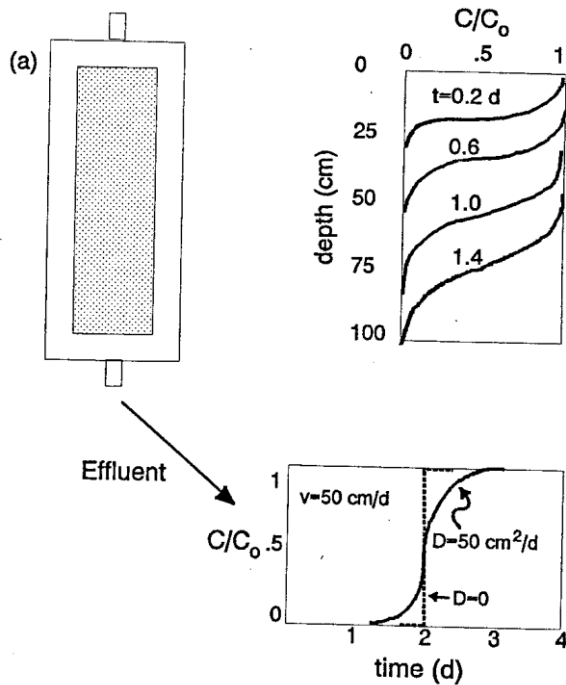


## Schematic of Transport Experiment

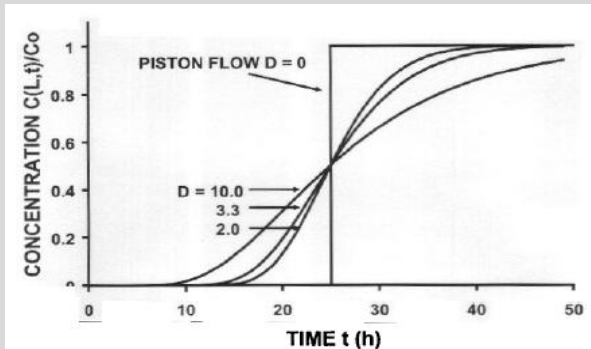
Breakthrough curves are often expressed in Pore Volumes,  $T$  (dimensionless time) and relative concentration (dimensionless)

$$T = \frac{\text{Volume Effluent}}{\text{Total Volume Solution in Column}}$$

$$T = \frac{qAt}{\theta AL} = \frac{qt}{\theta L} = \frac{vt}{L}$$



## Effect of Dispersion



The effect of dispersion on the breakthrough curve is to cause some early and late arrival of solute with respect to the breakthrough time. This deviation is due to diffusion and small-scale convection ahead of and behind the front moving at velocity  $V$  and becomes more pronounced as  $D$  becomes larger.

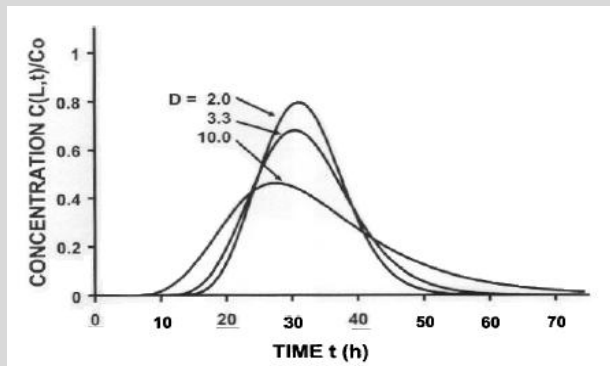
Outflow concentration versus time for a step change in solute input at  $t = 0$ . Column length is 50 cm and water velocity  $V = 2 \text{ cm h}^{-1}$ . Curves are for different values of  $D$  in  $\text{cm}^2\text{h}^{-1}$ .

$D = 0 \Rightarrow$  Piston Flow

$$C(L,t) = \frac{C_0}{2} \left[ \operatorname{erfc} \left( \frac{L-Vt}{\sqrt{4Dt}} \right) + \exp \left( \frac{VL}{D} \right) \operatorname{erfc} \left( \frac{L+Vt}{\sqrt{4Dt}} \right) \right]$$

$$\operatorname{erfc}(x) = \frac{2}{\sqrt{\pi}} \int_x^{\infty} \exp(-\xi^2) d\xi$$

## Effect of Dispersion

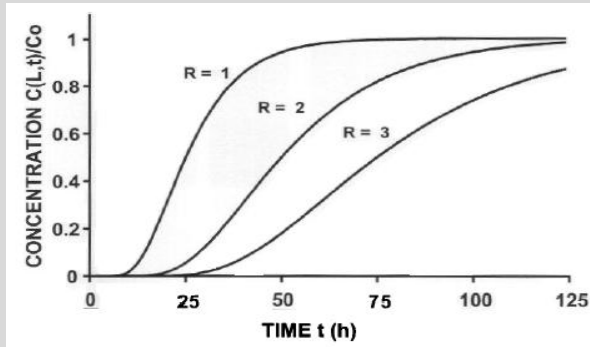


Note that as  $D$  becomes larger, the pulse spreads out more and the pulse height decreases.

Outflow concentration versus time for a narrow pulse of solute input at  $t = 0$ . Column length is 50 cm and water velocity  $V = 2 \text{ cm h}^{-1}$ . Curves are for different values of  $D$  in  $\text{cm}^2\text{h}^{-1}$ . ( $M_0$  – solute mass)

$$C(L,t) = \frac{M_0 L}{2J_w \sqrt{2Dt^3}} \operatorname{erfc} \left[ \frac{(L-Vt)^2}{2Dt} \right]$$

## Effect of Sorption



The breakthrough times increase as  $R$  increases, and the apparent amount of solute spreading of the outflow curves increases with  $R$  as well.

$K_d$  is the distribution or partition coefficient. For linear isotherm:

$$C_a = K_d C_l$$

$$R = 1 + \frac{\rho_b K_d}{\theta}$$

Outflow concentration versus time of chemicals with different retardation factors for a step change of solute input at  $t = 0$ . Column length is 50 cm and water velocity  $V = 2 \text{ cm h}^{-1}$ . Curves are for different values of  $R$  ( $K_d$ ).

## Effect of Sorption of solute to soil

$$R \frac{\partial c}{\partial t} = D \frac{\partial^2 c}{\partial x^2} - \bar{v} \frac{\partial c}{\partial x}$$

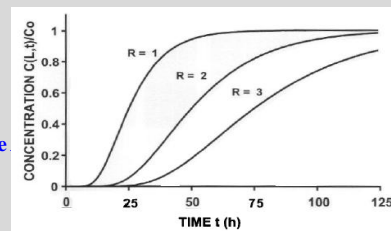
$$C_a = K_d C_l$$

$$R = 1 + \frac{\rho_b K_d}{\theta}$$

$$\frac{\partial c}{\partial t} = D_R \frac{\partial^2 c}{\partial z^2} - \bar{v}_R \frac{\partial c}{\partial z}$$

$$\bar{v}_R = \bar{v} / R$$

where





# Sorption

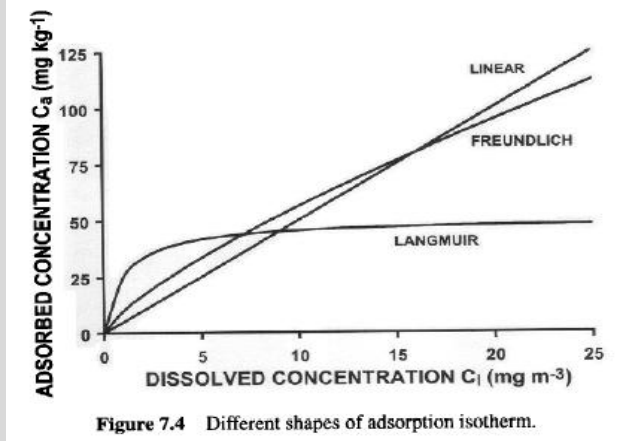
Adsorption isotherms:

Linear:

Freundlich:

Langmuir:

$k$  or  $K_d$  – distribution coefficient  
[L<sup>3</sup>/M], [m<sup>3</sup>/kg, cm<sup>3</sup>/g]



## One-Dimensional Equilibrium Transport

General Transport Equation; no production/decay

$$\frac{\partial(\theta c + \rho s)}{\partial t} = \frac{\partial}{\partial x} \left( \theta D \frac{\partial c}{\partial x} \right) - \frac{\partial qc}{\partial x}$$

Equilibrium Sorption ( $s = k c$ )

$$\frac{\partial(\theta R c)}{\partial t} = \frac{\partial}{\partial x} \left( \theta D \frac{\partial c}{\partial x} \right) - \frac{\partial qc}{\partial x}$$

$$R = 1 + \frac{\rho k}{\theta}$$

(Retardation Factor)

Steady-State Flow (homogeneous system; constant  $q$ ,  $\theta$ )

$$R \frac{\partial c}{\partial t} = D \frac{\partial^2 c}{\partial x^2} - v \frac{\partial c}{\partial x}$$

# *Initial and Boundary Conditions*

**Initial Condition:**  $c(x,0)=C_i$

**Inlet Boundary Conditions, at  $x=0$ :**

$$c(0,t) = C_0 \quad \left(-D \frac{\partial c}{\partial x} + v c\right)|_{x=0+} = v C_0$$

**Outlet Boundary Condition: Finite ( $x=L$ ) or semi-infinite ( $x \rightarrow \infty$ ) system**

$$\frac{\partial c}{\partial x}(L,t) = 0 \quad \frac{\partial c}{\partial x}(\infty,t) = 0$$

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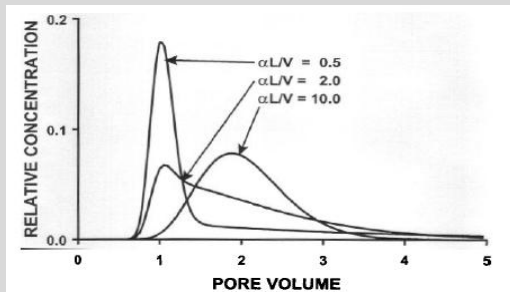
## One-Site Chemical Nonequilibrium

$$\frac{\partial(\theta c)}{\partial t} = \frac{\partial}{\partial z} \left( \theta D \frac{\partial c}{\partial z} - qc \right) - \alpha \rho (kc - s)$$

$$\frac{\partial s}{\partial t} = \alpha (kc - s)$$

## Effect of Nonequilibrium Sorption

When the time of contact is short compared to the time required for sorption, the assumption of instantaneous equilibrium between the adsorbed and dissolved chemical concentrations is not valid.



Outflow concentration versus pore volume for a chemical with a retardation factor  $R = 2$  as a function of the rate parameter  $\alpha$ .

For large values of  $\alpha$ , the sorption process reaches equilibrium during the time of contact, and the equilibrium sorption model may be used. When the rate parameter approaches zero, sorption is largely blocked, and the chemical moves like a mobile tracer. For intermediate values of  $\alpha$ , breakthrough begins earlier than with equilibrium sorption, and the outflow curve has a long tail.

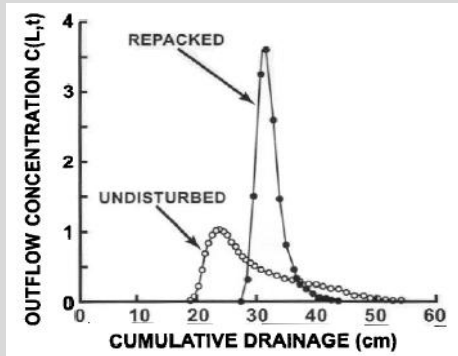
$$\rho_b \frac{\partial C_a}{\partial t} + \theta \frac{\partial C_l}{\partial t} = D_e \frac{\partial^2 C_l}{\partial z^2} - J_w \frac{\partial C_l}{\partial z}$$

$$\rho_b \frac{\partial C_a}{\partial t} = \alpha (K_d C_l - C_a)$$

$\alpha$  - a rate parameter that expresses the rate of mass transfer of chemical between the phases. Note that the rate of change of mass in the sorbed phase is assumed to be proportional to the deviation from equilibrium.

# Effect of Soil Structure

Aggregated soils have a continuous network of large pores through which water can flow rapidly, and a set of finer pores within the aggregates where water is present but largely stagnant. Soils with a massive structure may have fracture planes that channel virtually all of the water in the medium. In contrast, soils that have been packed uniformly have a distribution of pore sizes, and most of the water is either moving or in close contact with moving water. For such media, the uniform CDE generally provides a good description of the transport process.



Chloride arrives after less drainage in the undisturbed soil column than in the repacked one and has substantial delay in the arrival of the final portion of the applied pulse at the outflow end. For soils of this type, some of the solute can be transported much more rapidly than one would expect based on the total volumetric water content and the piston flow model.

# Mobile-Immobile Water Model

The mobile-immobile water model (MIM) divides the wetted pore space into two regions: a mobile water content ( $\theta_m$  through which water is flowing and an immobile water content ( $\theta_{im} = \theta - \theta_m$ ) which contains stagnant water. For this system, the solute concentration is divided into an average concentration  $C_m$  in the mobile region and a second concentration  $C_{im}$  in the immobile region. The solute is transported by a convection-dispersion process in the mobile region while mass is exchanged with the immobile region by a rate-limited diffusion process. For an inert, nonreactive solute:

$$\theta_m \frac{\partial C_m}{\partial t} + \theta_{im} \frac{\partial C_{im}}{\partial t} = D_e \frac{\partial^2 C_m}{\partial z^2} - J_w \frac{\partial C_m}{\partial z}$$

$$\theta_{im} \frac{\partial C_{im}}{\partial t} = \alpha (C_m - C_{im})$$

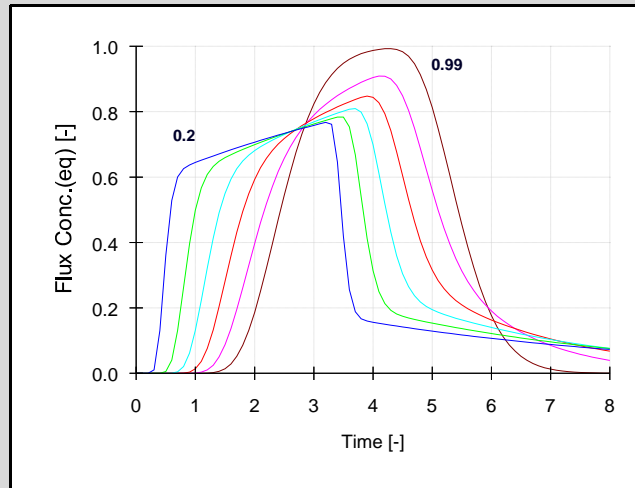
$\alpha$  - rate coefficient.

The rate of change of solute stored in the immobile region is proportional to the difference in concentration between the two regions. Thus, when solute is added to the soil as a pulse, it will first diffuse from the mobile to the stagnant region and then return to the mobile region after the pulse has passed through the system.

# *Mobile-Immobile Water Model*

## *Effect of mobile/immobile water ratio*

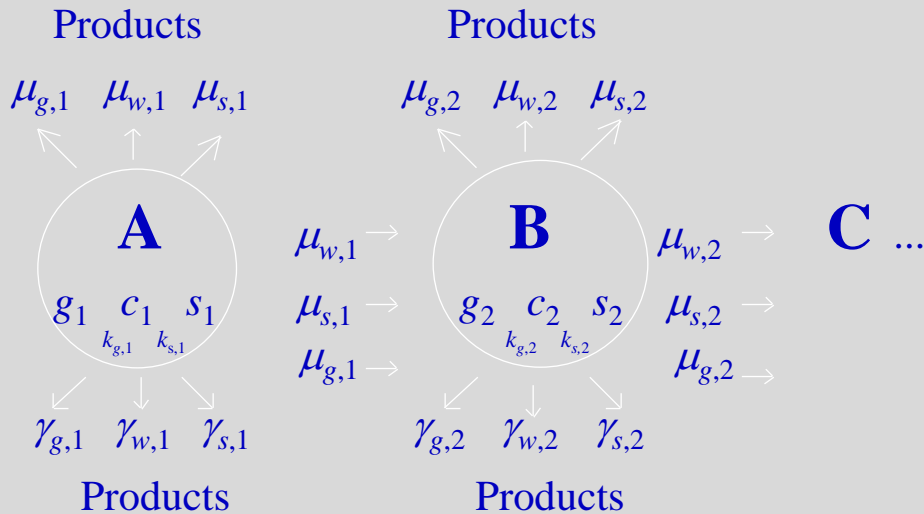
### Effect of mobile/immobile water ratio



## *Outline*

- ◆ Introduction: Water Flow and Solute Transport
- ◆ Solute Transport
  - Governing transport equations
  - Advection, diffusion, dispersion, reactions (ADE)
  - Solute Break-Through Curves, Boundary conditions
  - Nonequilibrium transport (chemical, physical)
  - **Decay chains**
  - Parameter estimation

## General structure of system of solutes:

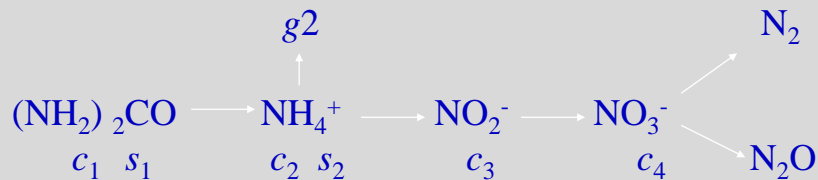


## Example sequential first-order chains:

### ◆ Radionuclides [van Genuchten, 1985]



### ◆ Nitrogen [Tillotson et al., 1980]



## Governing Solute Transport Equations

$$\frac{\partial(\theta c)}{\partial t} + \frac{\partial(\rho s)}{\partial t} = \frac{\partial}{\partial z} (\theta D \frac{\partial c}{\partial z} - qc) - \phi$$

$$\begin{aligned} \frac{\partial \theta c_k}{\partial t} + \frac{\partial \rho s_k}{\partial t} + \frac{\partial a g_k}{\partial t} = & \frac{\partial}{\partial z} (\theta D_k^w \frac{\partial c_k}{\partial z}) + \frac{\partial}{\partial z} (a D_k^g \frac{\partial g_k}{\partial z}) - \frac{\partial q c_k}{\partial z} - \\ & - (\mu_{w,k} + \mu_{w,k}') \theta c_k - (\mu_{s,k} + \mu_{s,k}') \rho s_k - (\mu_{g,k} + \mu_{g,k}') a g_k + \mu_{w,k-1}' \theta c_{k-1} + \\ & - \mu_{s,k-1}' \rho s_{k-1} - \mu_{g,k-1}' a g_{k-1} + \gamma_{w,k} \theta + \gamma_{s,k} \rho + \gamma_{g,k} a - S c_{r,k} \quad k \in (2, n_s) \end{aligned}$$

$w, s, g$  subscripts corresponding with the liquid, solid and gaseous phases, respectively

$c, s, g$  concentration in liquid, solid, and gaseous phase, respectively

## Governing Solute Transport Equations

- $q_i$  volumetric flux [ $LT^{-1}$ ]  
 $\rho$  soil bulk density [ $ML^{-3}$ ]  
 $a$  air content [ $L^3L^{-3}$ ]  
 $S$  sink term in the water flow equation [ $T^{-1}$ ]  
 $c_r$  concentration of the sink term [ $ML^{-3}$ ]  
 $D^w, D^g$  dispersion coefficients for the liquid and gaseous phase [ $L^2T^{-1}$ ], respectively  
 $k$  subscript representing the  $k$ th chain number  
 $\mu_w, \mu_s, \mu_g$  first-order rate constants for solutes in the liquid, solid, and gaseous phases [ $T^{-1}$ ], respectively  
 $\gamma_w, \gamma_s, \gamma_g$  zero-order rate constants for the liquid [ $ML^{-3}T^{-1}$ ], solid [ $T^{-1}$ ], and gaseous phases [ $ML^{-3}T^{-1}$ ], respectively  
 $\mu_w', \mu_s', \mu_g'$  first-order rate constants for solutes in the liquid, solid and gaseous phases [ $T^{-1}$ ], respectively; these rate constants provide connections between the individual chain species.  
 $n_s$  number of solutes involved in the chain reaction

## *Outline*

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  - **Parameter estimation**

## *Parameter Estimation with HYDRUS*

### **Parameter Estimation:**

- **Soil hydraulic parameters**
- **Solute transport and reaction parameters**
- **Heat transport parameters**

### **Sequence:**

- **Independently**
- **Simultaneously**
- **Sequentially**

### **Method:**

- **Marquardt-Levenberg optimization**



## *Objective Function for Inverse Problems*

$$\Phi(\mathbf{b}, \mathbf{q}, \mathbf{p}) = \sum_{j=1}^{m_q} v_j \sum_{i=1}^{n_{qj}} w_{i,j} [q_j^*(x, t_i) - q_j(x, t_i, \mathbf{b})]^2 +$$

$$\sum_{j=1}^{m_p} v_j \sum_{i=1}^{n_{pj}} w_{i,j} [p_j^*(\theta_i) - p_j(\theta_i, \mathbf{b})]^2 +$$

$$\sum_{j=1}^{n_b} \hat{v}_j [b_j^* - b_j]^2$$

- 1st term:** deviations between measured and calculated space-time variables
- 2nd term:** differences between independently measured,  $p_j^*$ , and predicted,  $p_j$ , soil hydraulic properties
- 3rd term:** penalty function for deviations between prior knowledge of the soil hydraulic parameters,  $b_j^*$ , and their final estimates,  $b_j$ .

## *Formulation of the Inverse Problem*

When the covariance matrix  $V$  is diagonal and all elements of matrix  $V_\beta$  are equal to zero, i.e., the measurement errors are noncorrelated and no prior information about the optimized parameters exists, the problem simplifies into the weighted least-squares problem

$$\Phi(\beta) = \sum_{i=1}^n w_i [C_i^* - C_i(\beta)]^2$$

$w_i$  - weight of a particular measured point

# Parameter Estimation

